



Studies on the Removal of Heavy Metals from Aqueous Solution Using Immobilized *Typha angustata L.*

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ABSTRACT

The sorption capacity of Immobilized *Typha angustata L.* for the removal of metal ions (Fe^{3+} , Pb^{2+} , Zn^{2+} , Cr^{3+} and Cd^{2+}) were studied using batch experiments. The residual metal ion concentrations were investigated using AAS. The influence of factors such as contact time, initial concentration, ionic strength pH and temperature were investigated. The metal ion sorption capacity by IBCC decreases with increasing ionic strength but increases with increasing pH values and initial metal ion concentration. The sorption rate for the metal ions was rapid for the first 30 minutes and equilibrium was established within 2 hrs. The maximum sorption capacity for Fe^{3+} , Pb^{2+} , Cd^{3+} , Cr^{2+} , and Zn^{2+} were found to be 98.65, 96.36, 90.57, 84.38 and 80.28 respectively. From the experimental work on the sorption capacity of IBCC sorbent, our result showed that the abundantly wasted *Typha angustata L.* could be turned into an environmentally friendly, and cost effective sorbent for the removal of heavy metal ions from waste water

1. Introduction

Due to the great amount of industrial activities, contamination of the environment by heavy metals has become a serious problem. Many of these toxic metals enter the environment through fossil fuel combustion as well as mining and smelting process (Wu et al., 2008). The natural process of metal transportation between the soil and water concentrates heavy metal contamination in the environment (Wuyep et al., 2007). Once in the environment, metals are difficult to remediate and can adversely impact human health. Even in low doses, metals which do not have distinguishable odour and colour characteristics, still pose a major threat. Current technologies for cleaning up heavy metals contaminated soils involve excavation and reburial in landfills. This process is expensive and usually reserved for small areas, and the metals are still not isolated from the environment. Contaminated waters are typically detoxified through the use of ion exchange and activated charcoal filters which are not only costly, but nonselective for heavy metals removal (Ogali et al., 2008). Due to the high cost of these methods, there is need for the development of a more cost effective method.

Bioremediation has emerged as an inexpensive solution. Many researchers have studied the use of live microbial systems for the purpose of remediating contaminated waters (Volesky, B. 2007). A few plant species have shown a remarkable resistance to heavy metals such as copper, zinc, and lead (Wuyep et al., 2007). This ability to grow in highly contaminated areas is thought to be due to the evolution of chemical functional groups which inhibit the toxicological effects of the heavy metals (Toti et al., 2008). Phyto-chelations and proteins within the plant may be produced in large concentration to bind the metals and reduce their harmful effects (Chamarthy et al., 2009). Therefore, plants that grow in contaminated areas should show a greater ability to recover heavy metals and may be a good source for naturally occurring biological compounds that have potential for contaminant remediation.

Although live biological systems work well for low concentrations, they cannot survive the high levels that are found in seriously contaminated areas and industrial influents. The use of non-living biomaterial containing metals-binding compounds would have the advantage of non requiring care and maintenance as well as being useful in remediating areas with high levels of contaminants that would otherwise kill live systems. Hence a plant such as *Typha angustata L.* which has the ability to tolerate heavy metals and adverse environments might contain natural metal binding compounds and could be a possible source for bioremediation of heavy metals in waste water (Alabuk et al., 2007)

The immobilization of *Typha angustata L.* biomasses involves the trapping of the biomass within the matrix of the polymer with sodium alginate and calcium chloride solutions to give a good biosorbent materials. However, biosorbents are hard enough to withstand the application of pressures, have water retention capacity, high and fast sorption uptake. Also because of immobilization, the biosorbents so obtained offer easy and convenient usage compared to free biomass, which is easily biodegradable (Volesky et al., 2007). One of the most valued properties of these modified product is their water absorbency and dehydration behaviour (Lee et al., 2009). They also exhibit suitable hydrophilicity and high swelling capacity. The target in this study is to immobilize the *Typha angustata L.* with sodium alginate with the objective of using the products as a sorbent for the removal of heavy metals in waste water.

Material and Methods

MATERIALS

Sodium alginate, Calcium chloride, Sodium hydroxide, Hydrochloric acid, Lead nitrate, Ferric chloride, Cadmium (II) nitrate, Chromium (III) nitrate, Zinc nitrate and were obtained from the British Drug House (BDH). The chemicals were of analytical grade and were used as supplied. *Typha angustata L.* was obtained from Perak, Malaysia

METHODS

Preparation of *Typha angustata L.* Calyx Sample The *Typha angustata L.* sample was sundried, decoated and milled into powder, then sieved through a 100 μm sieve screen to obtain fine powder and packaged separately in polyethylene bag for further use (wuyep et al., 2007).

Preparation of Sodium Alginate and Calcium Chloride Solutions

Sodium alginate was made by weighing 4.00g and making it up to 100 cm^3 mark with distilled water in a volumetric flask and left over night for complete dissolution to give 4% w/w solution. Calcium chloride (0.12M) was prepared by weighing 26.28g into 1 litre volumetric flask and making it up to the mark with distilled water (wuyep et al., 2007). **Synthesis of *Typha angustata L.* Immobilized**

25 cm^3 viscous layers of dissolved *Typha angustata L.* sample was mixed with 25 cm^3 of 4% stock solution of Sodium Alginate and stirred vigorously for even mixing in 250 cm^3 beaker. The mixture was then transferred into another beaker containing 30 cm^3 of 0.12M Calcium chloride solution. The reaction was allowed retention time of 1hr for complete precipitation. The precipitated blend solid of the sample was filtered and dried at room

temperature for 24hrs. The dried solid was stored in polyethylene bag for further use (Wuyep et al., 2007).

The above procedure was followed by mixing another set of blend compounds with 4% of the stock solution of Sodium alginate and *Typha angustata L.* in the ratio of 100:0, 95:5, 90:10, 85:15 and 80:20 which were designated as B1, B2, B3, B4 and B5 respectively. The dried solids were separately stored in polyethylene bags for further use (Wuyep et al., 2007).
Preparation of Synthetic waste water

The metal ions chosen for this study were Pb, Fe, Cr, Cd and Zn. The standards of these metal ion solutions were prepared from their salts by dissolving 1.60, 2.90, 7.70, 2.29 and 3.78 g of Lead nitrate, Ferric chloride, Chromium (III) nitrate, Cadmium (II) nitrate and Zinc nitrate in distilled water respectively and made up to 1 litre in a Volumetric flask that gave the stock solution of 1000 ppm of the metal ions. Serial dilution of 200 ppm of each metal ion solution was prepared with distilled water. The synthetic waste water was kept for further use.

Sorption experiment using Immobilized Typha angustata L.

The experiments were carried out in the batch mode for the measurement of adsorption capacities. From 200 ppm of each metal ion solution, 50 cm³ was taken into a 250cm³ conical flask and 0.2g of the IBCC was added corked with a rubber bung and shaken with a flask shaker for 2 hours at room temperature (30°C) at 150rpm. The separation of the adsorbents and solutions were carried out by filtration with whatman filter paper No 42 and the filtrates were stored in Sample cans for use. The residual metal ion Concentrations were determined using Atomic Absorption Spectrophotometer AAS Pyeunicam Model SP9 (Chamarthy et al., 2009; Osemeahon, 2003).

The percentage adsorption was calculated using the following equation
%Adsorption = $[(C_i - C_f) / C_i] \times 100$

Where C_i = Initial metal ion Concentration and C_f = Equilibrium metal ion Concentration (Barminas et al., 2005).

Determination of the effect of pH on Sorption Capacity.

The Sorption Capacity of the 0.2g of Immobilized *Typha angustata L.* with 50 ml of 200 ppm of various metal ion solutions at different pH values (1.0 to 6.0) were shaken for 2 hours using a flask shaker at 30°C. 1.0M HCl and 1.0M NaOH were used to adjust the pH of the solutions as the case may be. The residual metal ions were measured using Atomic Absorption Spectrophotometer AAS Pyeunicam Model SP9

(Chamarthy et al., 2009).

Determination of the effect of Contact time on Sorption Capacity.

The Sorption Capacity of IBCC for the various metal ions were determined. Set of samples consisting of 0.2g of the dried IBCC and 50ml of the metal ion solutions were introduced into 250ml conical flasks, corked with rubber bungs and climbed on a flask shaker. As the samples were undergoing agitation on the shaker, they were removed one after the other at a predetermined time interval ranging from 0.5 to 24 hours at 30°C. The solutions were filtered and analyzed for the residual metal ions using AAS (Chamarthy et al., 2009).
Determination of the effect of Initial metal ion Concentrations on Sorption Capacity.

The Sorption Capacity of various samples consisting of 50ml each of different metal ion concentrations ranging from 5ppm to 100ppm, each containing 0.2g of the dried IBCC were prepared and shaken for 2 hours on a flask shaker at 30°C. The solutions were filtered and analyzed for residual metal ions using AAS (Chamarthy et al., 2009).

Determination of the effect of temperature on Osemeahon et al., (2007); Wuyep et al., (2007) and Wu et al., (2008).

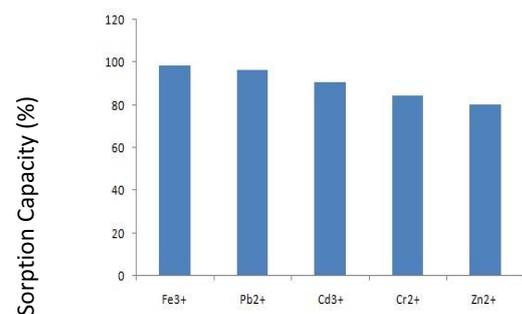


Figure 1: Sorption capacity of metal ions by *Typha angustata L.* The effect of pH on the maximum sorption capacity of the metal ions studied were 4 for Cd²⁺, 6 for Pb²⁺, and 5 for Fe³⁺, Cr³⁺, and Zn²⁺. The

pH of the sorption media is related to the metal sorption – mechanism of the surfaces from the water and reflected the nature of the physicochemical interaction of the ions in solution and nature of the sorption sites (Osemeahon et al., 2007). Therefore, it can be suggested that the differences in the pH optima for different metal ions could probably be due to the different solution chemistry of the metal ion (Lee et al., 2009).

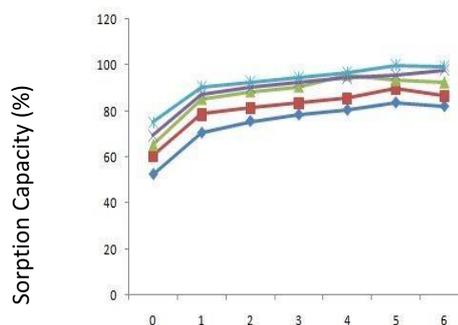


Figure 2: Effect of pH on sorption capacity of *Typha angustata L.* The pH influences the equilibrium metal ion uptake in aqueous solution. This happens because of counter reaction of the competing hydrogen ion as well as the chemistry of the active binding sites of the sorbent (Yenpeng et al., 2007).

Increasing the pH leads to precipitation of insoluble hydroxide or hydrated oxide, thereby lowering the metal ion availability for sorption. On the other hand, a decrease in pH results in the increase in the hydrogen ion concentration and hence possible competition for binding sites (Yenpeng et al., 2007).

Effect of Ionic Strength on Sorption of metal ions.

Industrial wastes and natural water often contain dissolved salts which compete with heavy metal ions for binding sites on the sorbent carbon chain (Osemeahon et al., 2008). There is the need to ascertain the influence of this factor on IBCC.

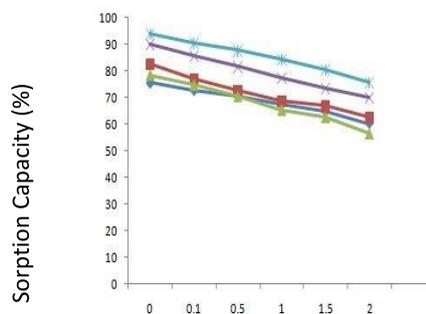


Fig. 3: Effect of ionic strength on sorption capacity of *Typha angustata L.*

The effect of the ionic strength on removal of the heavy metal ions from the aqueous solution by *Typha angustata L.* is represented in Figure 3. It can be observed that the sorbent showed a decrease in sorption capacity with increase in ionic strength for all the metal ions studied. The decrease in sorption capacity with increasing ionic strength is primarily attributed to the reduction in difference in the ionic osmotic pressure between the sorbent and the external solution (Lee et al., 2009; Osemeahon et al., 2008). The difference in osmotic pressure between the sorbent and the external solution decreases as the ionic strength of external solution increases. Therefore, the sorption of metal ion decreases when the ionic strength of external solution increases (Lee et al., 2009).

From this experiment, the coexistence of metal ions in same solution caused the interference with the performance of IBCC sorbent. Therefore the pretreatment of the effluent to remove the interfering ion appears to improve the efficiency of the sorbent (Kuyucak et al., 2008).

Effect of Contact Time on metal ions Sorption. The effect of contact time on sorption capacity of metal ions by *Typha angustata L.* is shown in Figure 4. Equilibrium was established after 2 hours of contact time for all the metal ions, but high sorption rates were observed during the first 30 minutes. Equilibrium time of 30 minutes was recorded for Pb²⁺ and two hours for the rest of the metal ions. The rapid sorption of metal ions may be attributed to highly porous structure of the sorbent which provide readily access for large surface area for the sorption of the metal ions to the binding

sites (Alabuk, et al., 2007).

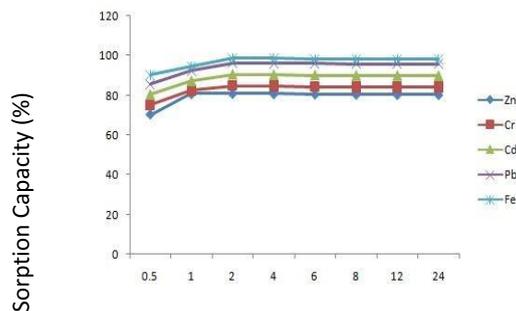


Figure 4: The effect of the contact time on sorption capacity of IBCC. This result demonstrates that the sorbent can compete favourably with other industrial sorbents in terms of kinetics, hence, equilibrium time is one of the important parameters for an economical wastewater treatment system. *Effect of Initial metal ion Concentration on Sorption.*

The result of metal ions sorption by the *Typha angustata* L. as a function of initial metal ion concentration is shown in Figure 5. It can be observed that the removal efficiency of the metal ions by the sorbent increases with the increase in initial concentration but leveled up however at certain point of the initial concentration.

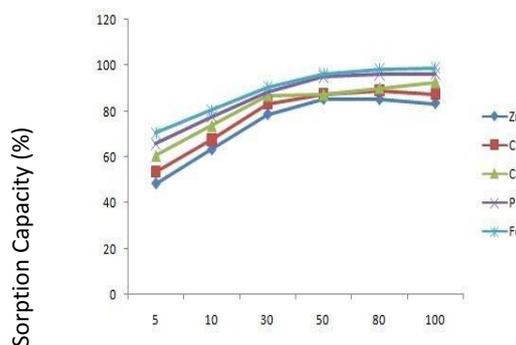


Figure 5: Effect of initial concentration on sorption capacity of IBCC.

This behaviour can be explained in term of increase in flux of the metal ions. The flux of the metal ions varies directly with the metal ion concentration and hence there should be an increase in flux with increase in initial concentration (Xiao et al., 2009; Osemeahon, 2007). The constant requirement observed for all metal ions explained the point of saturation. One of the problems of using conventional methods for the removal of the metal ions from aqueous systems is sometimes effectiveness, especially when the metal ion is present at a very low concentration (Toti et al., 2008). The behaviour of IBCC even at low concentration 5ppm shows that IBCC sorbent may be used in an attempt to improve the efficiency of metal ion removal from aqueous solution at low metal ion concentrations.

Effect of Temperature on heavy metal ions sorption. Figure 6 indicate the effect of temperature on the sorption of metal ions by *Typha angustata* L.. It can be seen that the sorption capacity decreases with an increase in temperature. The result of this work is in agreement with previous reports (Volesky, B. 2007; Lee et al., 2009; Ju et al., 2008; Osemeahon et al., 2003; Mahavi et al., 2008 and Ogali et al., 2008). From the graph, it can be seen that the sorption capacity of the metal ions decreases sharply from 30°C to 50°C. The decrease of the polymer matrix performance with increase in temperature may be due to the following factors:

- i. The inherent molecular structure of the polymer matrix formed (Ogali et al., 2008).
- ii. Concentration of the immobilized sample molecules with increase in temperature to give a more compacted form of the polymer matrix which causes the pore to be narrower and suction sites be hidden or inaccessible to the water molecules (Barminas and Osemeahon, 2005).

- iii. The sorption mechanism may be a physical adsorption Physical adsorption are normally exothermic, thus the extent of adsorption generally increases with decreasing temperature (Ju et al., 2008)

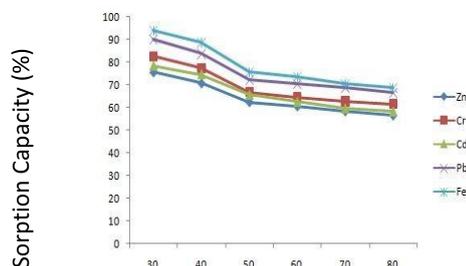


Figure 6: Effect of temperature on sorption capacity of *Typha angustata* L.

Conclusion

The sorption of Zn, Cr, Cd, Pb and Fe ions by *Typha angustata* L. has been shown to be dependent upon pH, with best sorption occurring between pH 5 and 6. This effect in pH suggests that the sorption mechanism may be an ion-exchange type process. Also, the sorption mechanism for these metals is a stable, rapid process which implies that the sorption is taking place on the cell wall surface of the *Typha angustata* L.. The experiments have demonstrated that *Typha angustata* L. possessed the ability to bind appreciable amounts of Zn, Cr, Cd, Pb and Fe ions as compared with other biosorbents. This ability to remove heavy metals from solution indicates the tremendous potential that the *Typha angustata* L. could have for cleaning the environment and industrial waste effluents from toxic metal ions.

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